

# Greenhouse gas mitigation potentials in the livestock sector

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**The livestock sector supports about 1.3 billion producers and retailers, and contributes 40–50% of agricultural GDP. We estimated that between 1995 and 2005, the livestock sector was responsible for greenhouse gas emissions of 5.6–7.5GtCO<sub>2</sub>e yr<sup>-1</sup>. Livestock accounts for up to half of the technical mitigation potential of the agriculture, forestry and land-use sectors, through management options that sustainably intensify livestock production, promote carbon sequestration in rangelands and reduce emissions from manures, and through reductions in the demand for livestock products. The economic potential of these management alternatives is less than 10% of what is technically possible because of adoption constraints, costs and numerous trade-offs. The mitigation potential of reductions in livestock product consumption is large, but their economic potential is unknown at present. More research and investment are needed to increase the affordability and adoption of mitigation practices, to moderate consumption of livestock products where appropriate, and to avoid negative impacts on livelihoods, economic activities and the environment.**

The livestock sector is large. Twenty billion animals make use of 30% of the terrestrial land area for grazing, one-third of global cropland area is devoted to producing animal feed<sup>1</sup> and 32% of freshwater is used to provide direct livelihood and economic benefits to at least 1.3 billion producers and retailers<sup>2,3</sup>. As an economic activity, livestock contributes up to 50% of agricultural GDP globally<sup>4</sup>. The livestock sector is also very dynamic. Global per capita consumption of livestock products has more than doubled in the past 40 years<sup>4</sup>. Increasing human population, incomes and urbanization are projected to drive increases in the consumption of milk and meat over the next 20 years, at least at previously observed rates<sup>1,5</sup>, with most of the growth projected to occur in the developing world. In response to these demand trends, the sector has managed to significantly increase production. Beef and milk production have more than doubled over the past 40 years and monogastric production (pigs and poultry) has grown in places by a factor of five or more<sup>2</sup>. Intensification of production, in terms of increased livestock and/or crop productivity, has played a pivotal role in raising the output per unit of land and animal<sup>1</sup>. For example, in the USA, 60% more milk is produced now than in the 1940s with about 80% fewer cows<sup>6</sup>. Although intensification has occurred in some regions, agricultural land expansion has also been an important component of production growth in places such as Africa and Latin America. These trends, if continued, could drive significant increases in greenhouse gas (GHG) emissions, deforestation, loss of biodiversity and other negative impacts on the environment<sup>7</sup>.

Here we review the mitigation potential of a number of field-tested management options for mitigating GHG emissions in livestock production. Our Review incorporates new supply-side information, such as changes in the structure of livestock production systems, with information about how policies that reduce demand for animal protein might contribute to emissions reductions. This synthesis of supply- and demand-side mitigation potentials is central to understanding the ways in which the components of our food systems interact and how livestock emissions could evolve and be managed in the future. We focus on biogenic emission sources in livestock systems, so we do not address CO<sub>2</sub> from energy use.

Mitigation potentials were estimated for the following: (i) technical and management interventions; (ii) intensification and the associated structural changes of livestock systems; and (iii) moderation of demand for livestock products.

We estimate that these options have the technical potential to mitigate a substantial proportion of emissions from livestock. However, their economic mitigation potential may be far smaller due to adoption barriers and costs of the technical options, underinvestment in the livestock sector and a lack of effective policies or political will for promoting healthy levels of consumption of livestock products in the diets of developed and developing country populations. We conclude with a discussion on research needs for improving the feasibility and adoption of mitigation options in livestock systems to lower GHG emissions without hampering rural economies and livelihoods.

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**Table 1 | Current global greenhouse gas emissions from livestock (~1995–2005).**

Emissions source	Emissions (GtCO <sub>2</sub> e)	Reference
Feed N <sub>2</sub> O	1.3–2.0 <sup>‡</sup>	9,13,15,16–18
Feed CO <sub>2</sub> (LUC excluded)	0.92	15,17,18
Feed CO <sub>2</sub> (LUC)	0.23	15,17,18
Pasture expansion CO <sub>2</sub> LUC	0.43	15,17,18
Feed CH <sub>4</sub> rice	0.03	15,17,18
Enteric CH <sub>4</sub> *	1.6–2.7	9–13,15,17
Manure CH <sub>4</sub> *	0.2–0.4	9–13,15,17,18
Manure N <sub>2</sub> O*	0.2–0.5	9–13,15–18
Direct energy CO <sub>2</sub>	0.11	15,17,18
Embedded energy CO <sub>2</sub>	0.02	15,17,18
Post-farm gate CO <sub>2</sub>	0.023	15,17,18
Non-CO <sub>2</sub> emissions* (IPCC guidelines)	2.0–3.6	This Review
Total emissions (LCA approach) <sup>†</sup>	5.6–7.5	This Review

\*Livestock emissions according to IPCC emissions guidelines<sup>8</sup>. <sup>‡</sup>Range estimated using information from global analyses for key emissions source categories. LCA as implemented by FAO<sup>15</sup>. <sup>†</sup>Includes N<sub>2</sub>O emissions from manures applied to pastures, and from fertilizers to croplands for both feed and pasture. Emissions from manure applied to pastures ranges from 0.42–0.95 GtCO<sub>2</sub>e. LUC, land-use change.

## GHG emissions from livestock

A key starting point for assessing the mitigation potential of the livestock sector is understanding the baseline levels of emissions associated with different livestock products, the key contributing greenhouse gases, their main sources and the regional differences in livestock production systems. We present these below.

**Current emissions.** Several global estimates of GHG emissions from livestock have been undertaken (Table 1). Here we classify studies as either following IPCC emissions guidelines<sup>8</sup>, which include direct non-CO<sub>2</sub> emissions of methane (CH<sub>4</sub>; enteric and manure) and nitrous oxide (N<sub>2</sub>O, manure management)<sup>9–13</sup>, or using life-cycle analysis (LCA)<sup>14,15</sup>, which includes extra sources in the supply chain; from conception to retail, emissions arise from feed production and animal rearing as well as from the processing and transportation of livestock commodities to markets. Further emissions occur after sale associated with transportation, storage, cooking and consumption or possible disposal. Some of these LCA emissions are reported in the GHG inventories of other sectors (that is, fuels to transport products in the transport sector, energy used in processing in industry sector).

We estimate that total emissions from livestock from 1995 to 2005 were between 5.6 and 7.5 GtCO<sub>2</sub>e yr<sup>-1</sup> (Table 1). The most important sources of emissions were enteric CH<sub>4</sub> (E<sub>CH<sub>4</sub></sub>; 1.6–2.7 GtCO<sub>2</sub>e yr<sup>-1</sup>; refs 9–13,15), N<sub>2</sub>O emissions associated with feed production (1.3–2.0 GtCO<sub>2</sub>e yr<sup>-1</sup>; ref. 15) and land use for animal feed and pastures, including change in land use (~1.6 GtCO<sub>2</sub>e yr<sup>-1</sup>; ref. 15). Figure 1 shows the spatial distribution of global livestock emissions<sup>13</sup>.

The level of disaggregation of global livestock emissions differs considerably between studies (see Supplementary Information). Some estimates are based primarily on Tier 1 approaches (using default global or regional emission factors)<sup>10–12,16</sup>, with Tier 2 approaches (using estimated regional or local emission factors) sometimes being used for enteric fermentation<sup>9</sup>. Some studies disaggregate emissions by country and region, species, production system and by product (milk, meat)<sup>13,15</sup>. FAO estimates<sup>15,17,18</sup> use Tier 2 methods for the IPCC emissions categories and LCA methods for the other sources. Herrero *et al.*<sup>13</sup> use Tier 3 methods (application of a rumen kinetics model) for E<sub>CH<sub>4</sub></sub> and Tier 2 methods for the other source categories.

Cattle production systems dominate the sector's emissions (64–78% depending on the study<sup>13,15,17</sup>). FAO studies<sup>17,18</sup> using LCA estimated cattle emissions from all sources to be about 4.6 GtCO<sub>2</sub>e yr<sup>-1</sup>, of which 2.5 GtCO<sub>2</sub>e yr<sup>-1</sup> were from beef cattle and 2.1 GtCO<sub>2</sub>e yr<sup>-1</sup> from the dairy cattle herd (producing both milk and meat). The other species have much lower and similar levels of emissions: pig (0.7 GtCO<sub>2</sub>e yr<sup>-1</sup>), poultry (0.7 GtCO<sub>2</sub>e yr<sup>-1</sup>), buffalo (0.6 GtCO<sub>2</sub>e yr<sup>-1</sup>) and small ruminants (0.5 GtCO<sub>2</sub>e yr<sup>-1</sup>).

Soils are the dominant source within the global atmospheric budget of N<sub>2</sub>O. Emissions of N<sub>2</sub>O due to agriculture activities are estimated at 2.8–6.2 TgN<sub>2</sub>O yr<sup>-1</sup>, 20–40% of all sources<sup>19–21</sup>, of which emissions associated with feed production may account for 1.3–2.0 GtCO<sub>2</sub>e yr<sup>-1</sup> (Table 1). Nitrous oxide emissions are directly linked to the use of synthetic and organic fertilizers for food and feed production and to livestock manure management and urine excretion to grazed grasslands.

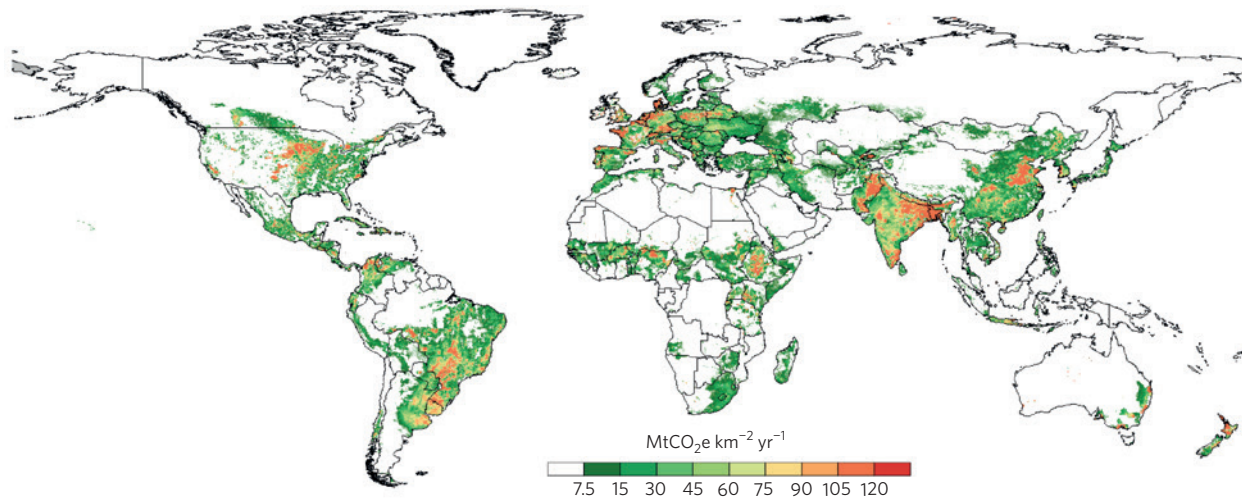
The developing world (non-Annex I countries; see Supplementary Information) contributes 70% of non-CO<sub>2</sub> emissions from ruminants and 53% from monogastrics<sup>13</sup>, and this share is expected to grow as livestock production increases to meet demand growth in the developing world. Mixed crop–livestock systems dominate livestock emissions (58% of total emissions) largely because of their prevalence, whereas grazing-based systems contribute 19% (ref. 13). Industrial and other systems comprise the rest.

Taking an aggregate view of the sector, and using all LCA sources of emissions, animal feed production accounts for about 45% of the sector's emissions, with about half of these emissions related to fertilization of feed crops and pastures (manure and fertilizer included)<sup>15</sup>. The remaining animal feed emissions are shared between energy use and land use. Enteric fermentation contributes about 40% of total emissions, followed by manure storage and processing (~10% of emissions)<sup>17,18</sup>.

Direct energy consumption on animal farms, energy consumption embedded in farm buildings and equipment and post-farm gate emissions account for less than 5% of the sector's emissions. However, when added to the energy consumption related to animal feed production, energy accounts for about 20% of the sector's emissions<sup>15</sup>.

Methane (CH<sub>4</sub>) accounts for 43% of emissions, and the remaining part is almost equally shared between N<sub>2</sub>O (29%) and CO<sub>2</sub> (27%). These estimates exclude carbon sequestered in grazing land (rangeland and pastures)<sup>16</sup>. There is some consensus on the magnitude of E<sub>CH<sub>4</sub></sub> emissions, irrespective of the approach used (mean 2.0 GtCO<sub>2</sub>e yr<sup>-1</sup>, coefficient of variation (CV) = 18%). Methane and N<sub>2</sub>O emissions from manure management, although smaller in magnitude, are more uncertain at the global level (mean 0.28 GtCO<sub>2</sub>e yr<sup>-1</sup>, CV = 27%; mean = 0.29 GtCO<sub>2</sub>e yr<sup>-1</sup>, CV = 46%). Comparable uncertainties (11–145%) for CH<sub>4</sub> emissions from manure management for several European countries have also been reported<sup>22–24</sup>, and those for European CH<sub>4</sub> emissions from enteric fermentation agree with the global level estimates (6–40%)<sup>23,25,26</sup>. Reported national N<sub>2</sub>O emissions from manure management (storage only) are uncertain in the range of 21–414%, whereas direct and indirect N<sub>2</sub>O emissions from agricultural land due to fertilizer application or soil N<sub>2</sub>O emissions from grazing animals (for example, urine patches) have an uncertainty of 57–424% (mean value: 156) at the national level<sup>23</sup>.

**Baseline projections.** Estimates of the emissions associated with the projected growth of the livestock sector to 2050 suggest that E<sub>CH<sub>4</sub></sub>, CH<sub>4</sub> from manure management and N<sub>2</sub>O from manure management are likely to grow at rates between 0.9–5%, 0.9–4%, 1.2–3% per year, respectively<sup>10,11,16,27–29</sup>. The ranges reflect different scenarios and assumptions about growth in demand for livestock products, animal numbers and type and productivity growth in livestock systems. A continuation of existing trends would lead to increases in



**Figure 1 | GHG emissions from global livestock for 1995–2005.** Using data from 2005, we have updated the version of this map used in ref. 13.

livestock emissions of between 1–1.5% per year across all sources other than land-use change (Fig. 2)<sup>10,29</sup>. Although not only attributable to livestock, emissions from deforestation over the same period are projected to grow at a slower rate of 0.8% yr<sup>-1</sup> (ref. 29). Cropland area expansion is growing at a faster rate than pasture expansion, primarily due to the accelerated growth of pig and poultry production (>5% globally).

**Emissions intensities in livestock systems.** The global non-CO<sub>2</sub> emissions intensity of livestock products is estimated at 44 kgCO<sub>2</sub>e kg<sup>-1</sup> protein, with a large range between 9–500 kgCO<sub>2</sub> kg<sup>-1</sup> (ref. 13). The range reflects differences between livestock products, with monogastrics (pigs and poultry) at the lower end of the range, followed by milk and ruminant meats<sup>13,15,30,31</sup>. The developed world has high absolute emissions but significantly lower emissions intensities than the developing world due to improved livestock diets, genetics, health and management practices. These practices reduce CH<sub>4</sub> emissions intensities and CO<sub>2</sub> emissions intensities due to lower land-use requirements. Many parts of the developing world have high emissions from livestock, which are produced at high emissions intensities due to low productivity and large numbers of animals (for example, parts of Africa and Latin America)<sup>13</sup>.

### Mitigation options and their technical potential

The mitigation options presented below can be targeted towards the supply of livestock products. These include technical and management interventions, and practices for increasing crop and livestock productivity. Another group of options could target reductions in the consumption of livestock products.

### Technical and management interventions

Several mitigation practices could be implemented within the livestock sector (emissions from crop production for livestock feed are not considered here) (Fig. 3). Animal-based GHG mitigation options for livestock can be categorized as targeting E<sub>CH<sub>4</sub></sub>, manure storage and application or deposition and animal management. Hristov *et al.*<sup>32</sup> provides a comprehensive description of these. We estimate that these practices could help mitigate between 0.01–0.5 GtCO<sub>2</sub>e yr<sup>-1</sup>. In ruminant production systems, E<sub>CH<sub>4</sub></sub> emissions usually comprise the largest proportion of GHG emissions and have been the main focus for animal-based mitigation research<sup>33–35</sup>.

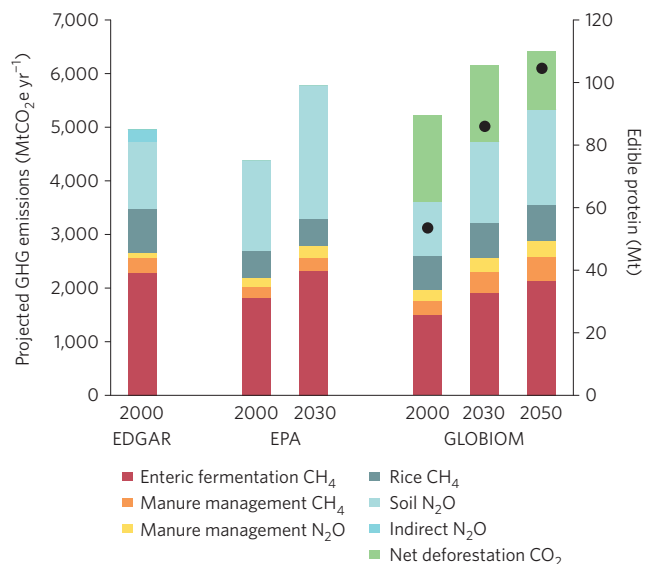
**Use of feed additives.** Several chemical compounds, such as alternative electron receptors, ionophoric antibiotics, enzymes and probiotic cultures, have been tested for their ability to decrease CH<sub>4</sub>

emissions, mainly in short-term experiments<sup>32</sup>. Their effects are often much reduced in the long term due to adaptation of the rumen microbial ecosystem. In addition, there may be as yet unknown environmental concerns associated with some of these compounds, and issues around acceptance by the public may severely constrain widespread adoption. Nevertheless, recent experiments with new CH<sub>4</sub> inhibitors show promise for the future<sup>36</sup>.

**Improved feed digestibility.** A well-studied E<sub>CH<sub>4</sub></sub> mitigation option for ruminants is the provision of forages of higher digestibility. This is unlikely to yield many benefits in highly developed animal production systems but has considerable potential in developing countries<sup>37</sup>. Another well-studied option for decreasing E<sub>CH<sub>4</sub></sub> emissions and increasing efficiency is inclusion of energy-dense feeds (for example, cereal grains) in the ration, with the greatest potential in production systems that utilize little or no grain to feed animals, as is common in many parts of the world. Widespread adoption of this practice depends on economic feasibility though improving the nutritive value of low-quality feeds but could have a considerable effect on herd productivity while keeping E<sub>CH<sub>4</sub></sub> emissions constant<sup>30</sup>. To maximize the benefits of improved feed quality, reductions in animal numbers need to be part of the strategy. Fewer better-fed animals could reduce pressure on land and other resources, but greater economic return from more efficient systems may encourage farmers to keep more livestock<sup>37</sup>. Our estimated technical mitigation potential of this practice is 0.68 GtCO<sub>2</sub>e yr<sup>-1</sup>, when a 10% increase in digestibility of the basal diet is considered and is widely applied throughout the developing world. The economic mitigation potential, however, is closer to 0.12–0.15 GtCO<sub>2</sub>e yr<sup>-1</sup> when considering the low adoption rates (20–25%) of improved feeding practices in the developing world over the last 20 years<sup>37,38</sup>.

Forages with a high concentration of plant secondary metabolites such as tannins have also been shown to decrease E<sub>CH<sub>4</sub></sub>, although results have been inconsistent. Inclusion of lipids or high-oil by-product feeds, such as distiller's grains, when available, may be an economically feasible mitigation practice<sup>39</sup>. Algae is another option that merits more investigation, as its high quality could lead to significant yield improvements and reduced land use, but current economic costs and logistics prevent its widespread uptake<sup>40</sup>.

**Manure management.** Emissions can be reduced by decreasing N losses, for example, by storing manure or slurries appropriately, thereby minimizing losses due to volatilization or runoff<sup>41</sup>. Often simple measures can be taken to avoid nutrient losses to the environment such as compacting and covering farmyard manure<sup>42,43</sup>.



**Figure 2 | Baseline projections of GHG emissions for the main IPCC source categories for the entire agricultural sector.** The baseline projection represents a continuation of the current livestock product demand trends (black dots, converted to edible animal protein for all livestock products). EDGAR data are from ref. 9, EPA data from ref. 27 and GLOBIOM data from ref. 29.

Slurry may also be anaerobically digested before its application, which affects the organic matter content and the concentrations of volatile solids but N content is minimally affected. However, evidence for anaerobic digestion reducing field-scale N<sub>2</sub>O emissions is mixed<sup>44,45</sup>. For most livestock systems worldwide, there is limited opportunity for manure management, treatment or storage<sup>41</sup>; most excretion happens in the field and collection for fuel or fertility amendment occurs after it is dry and CH<sub>4</sub> emissions are negligible. The greatest mitigation potential possibly derives from application of manures to the field; the N<sub>2</sub>O mitigation potential ranges from 0.01–0.075 GtCO<sub>2</sub>e yr<sup>-1</sup> (ref. 41) depending on timing and form of application<sup>46,47</sup>. Even if N<sub>2</sub>O emissions increase following N application, the emissions per unit of product, which is the most important agronomic criterion<sup>48</sup>, is likely to be reduced if manures are applied to match plant N demand at times that avoid heavy rains<sup>49</sup>. Other options for reducing N<sub>2</sub>O include the use of nitrification inhibitors<sup>50</sup>, which have been successfully tested for suitability in reducing N<sub>2</sub>O emissions from cropland and grassland in various climates<sup>50–52</sup>.

**Soil carbon sequestration in grasslands.** Grazing-land management practices that affect species composition, forage consumption, nutrient and water inputs and fire can impact soil carbon stocks<sup>53</sup>. Excessive removal of aboveground biomass, continuous grazing at suboptimal stocking rates and other poor grazing management practices that result in a mismatch between forage supply and animal demands have led to the depletion of soil carbon stocks<sup>53,54</sup>. Much of the world's grazing land is still under pressure to produce more livestock through expansion and more intensive grazing, particularly in Africa<sup>55</sup>. However, good grassland management can potentially reverse historical soil carbon losses and sequester substantial amounts of carbon in grazing-land soils (Fig. 4a). Much of this sequestration potential may be economically feasible, because it can be realized by implementing practices that enhance forage production<sup>41</sup>. Changes in grazing management — increasing or reducing the forage consumption rate to maximize forage production — could lead to annual sequestration of up to 150 MtCO<sub>2</sub>e yr<sup>-1</sup> in the world's grazing lands<sup>56</sup> (Fig. 4b). Much of this potential (81%, approximately 120 MtCO<sub>2</sub>e yr<sup>-1</sup>) is in developing countries,

occurring in areas where production is predicted to increase following a period of de-stocking, that is, in areas where primary production can recover from overgrazing<sup>56</sup>. The use of legumes in pastures has been estimated to sequester 200 MtCO<sub>2</sub>e yr<sup>-1</sup> globally, though this could increase soil N<sub>2</sub>O emissions by 60 MtCO<sub>2</sub>e yr<sup>-1</sup>, offsetting 28% of the soil carbon sequestration benefits. Although recent evidence suggests that modest amounts of carbon can be sequestered, about half of the global net mitigation potential of this option is in developing countries<sup>56</sup>. Carbon sequestration in grazing lands should perhaps be considered a co-benefit of improving productivity and ecosystems services<sup>57</sup>, rather than a primary objective for managing grazing land ecosystems.

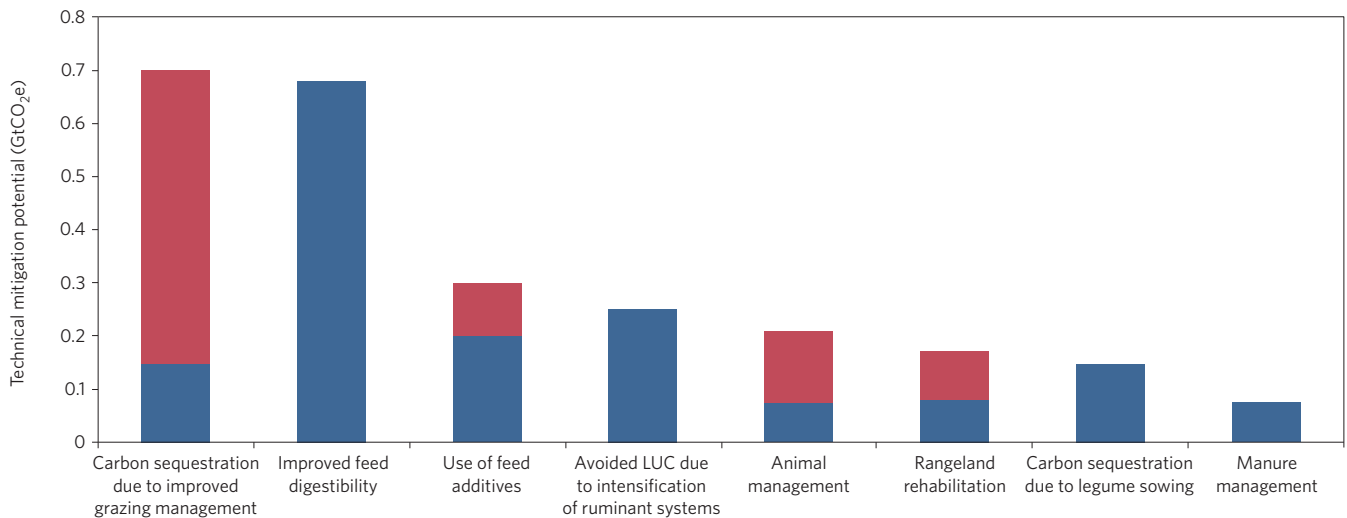
**Increased livestock and crop/pasture productivity**

Practices that increase livestock, crop and pasture productivity have received significant attention in the past years, due to their multiple mitigation benefits. They can improve the GHG emission intensities of different livestock products, but they can also have indirect benefits associated with land use sparing, and could promote structural changes in the livestock sector. This section briefly describes these mitigation benefits.

**Animal productivity and health.** Improving the genetic potential of animals for production, their reproductive performance, health and liveweight gain rates are among the most effective approaches for reducing GHG emissions per unit of product<sup>32,58</sup>. In subsistence agricultural systems, reduction of herd size by increased reproduction rates increases feed availability as well as the productivity of individual animals and the total herd, thus lowering E<sub>CH<sub>4</sub></sub> and overall GHG emissions per unit of product. Reducing the age at slaughter by increasing liveweight gain rates significantly decreases GHG emissions per unit of product in beef and other meat production systems. Improved animal health and reduced mortality and morbidity can increase herd productivity and reduce emissions intensity in all livestock systems. Adoption of modern reproductive management technologies, targeting increased conception rates, increased fecundity (in swine and small ruminants) and reduced embryo loss also provide a significant opportunity to reduce GHG emissions from the livestock sector, with appropriate attention to animal welfare considerations. We estimated that these improved animal management practices could reduce emissions in the livestock sector by 0.2 GtCO<sub>2</sub>e yr<sup>-1</sup> by 2050.

**Avoided deforestation due to intensification.** A recent review by Smith<sup>59</sup> indicates that this involves addressing the many unsustainable practices already manifest in the global food system. Many options for sustainable intensification exist, ranging from adoption of new technology to improving the efficiency of current food production. At the high-tech end are options such as the genetic modification of living organisms and the use of cloned livestock and nanotechnology<sup>60–62</sup>. By 2050 it may be possible to manipulate the traits controlled by many genes and confer desirable traits (such as improved nitrogen and water-use efficiency in crops) with improved productive characteristics<sup>61</sup>. But the future role of genetic manipulation in sustainable intensification is heavily dependent on the softening of public opposition, which is widespread in places.

It may be possible in some regions to rebalance the distribution of inputs to optimize production and close yield gaps<sup>63,64</sup>. The benefits and impacts of irrigation and the water needed for crop production vary greatly across the globe, but addressing these imbalances could bring yields to within 95% of their current potential for 16 important food and feed crops, and adding 2.3 billion tonnes (5 × 10<sup>15</sup> kcal) of new production (a 58% increase)<sup>63</sup>. Closing the yield gap of the same crops to 75% of their potential could increase global production by 1.1 billion tonnes (2.8 × 10<sup>15</sup> kcal), an increase of 28% (ref. 63).



**Figure 3 | Technical mitigation potentials of supply-side options for reducing emissions from the livestock sector.** Red represents the range for each practice, where available. The range of mitigation potentials for carbon sequestration due to improved grazing management is defined by refs 41,56. The mitigation potentials for 10% improved digestibility in all ruminants in the developing world are shown, obtained by up-scaling values from ref. 37. Direct application of this option to developed country situations was assumed to be too small to be considered. Full adoption of the practice across all ruminants was assumed to obtain the technical mitigation potential for this practice, hence no range is presented. The mitigation potentials for use of feed additives are from ref. 32, and include inhibitors, ionophores, electron receptors, enzymes, plant bioactive compounds, lipids and manipulation of rumen microflora. These potentials are applied to breeding herds of cattle globally with effects on  $E_{\text{CH}_4}$  as described in ref. 32. The mitigation potentials of avoided land-use change from transitions from grazing to mixed crop-livestock systems are from ref. 79. The mitigation potentials for animal management practices such as improved health and reduced mortality are from ref. 32, and the effects applied as for feed additives. Rangeland rehabilitation mitigation potentials are from ref. 56. Manure management mitigation potentials are from ref. 41.

Crop yield improvements will play a critical role in future land-use dynamics<sup>65</sup> and livestock systems<sup>32</sup>. They will determine the requirements for extra cropland and strongly affect grassland expansion<sup>32</sup>. Compared with yield stagnation, maintaining past trends in crop yield growth would save 290 Mha of cropland and avoid further expansion of about 120 Mha of grassland by 2030<sup>32</sup>. This is caused by increasing crop yields, which lower crop prices, leading to shifts in ruminant production from grass-based systems to more intensive systems with forage-based diets supplemented with grains; and relocation of livestock production to the most cost-effective regions. Hristov *et al.*<sup>32</sup> found that GHG emissions decreased by more than 2 GtCO<sub>2</sub>e yr<sup>-1</sup> when crop yields grew according to past trends, compared with yield stagnation. About 90% of the emissions reduction came from avoided land-use changes partly associated with livestock (0.25 GtCO<sub>2</sub>e yr<sup>-1</sup>). Emissions associated with feed production were also reduced due to improved productivity. Productivity increases based solely on higher fertilizer rates would reduce the overall positive balance through increased crop N<sub>2</sub>O emissions<sup>66</sup>.

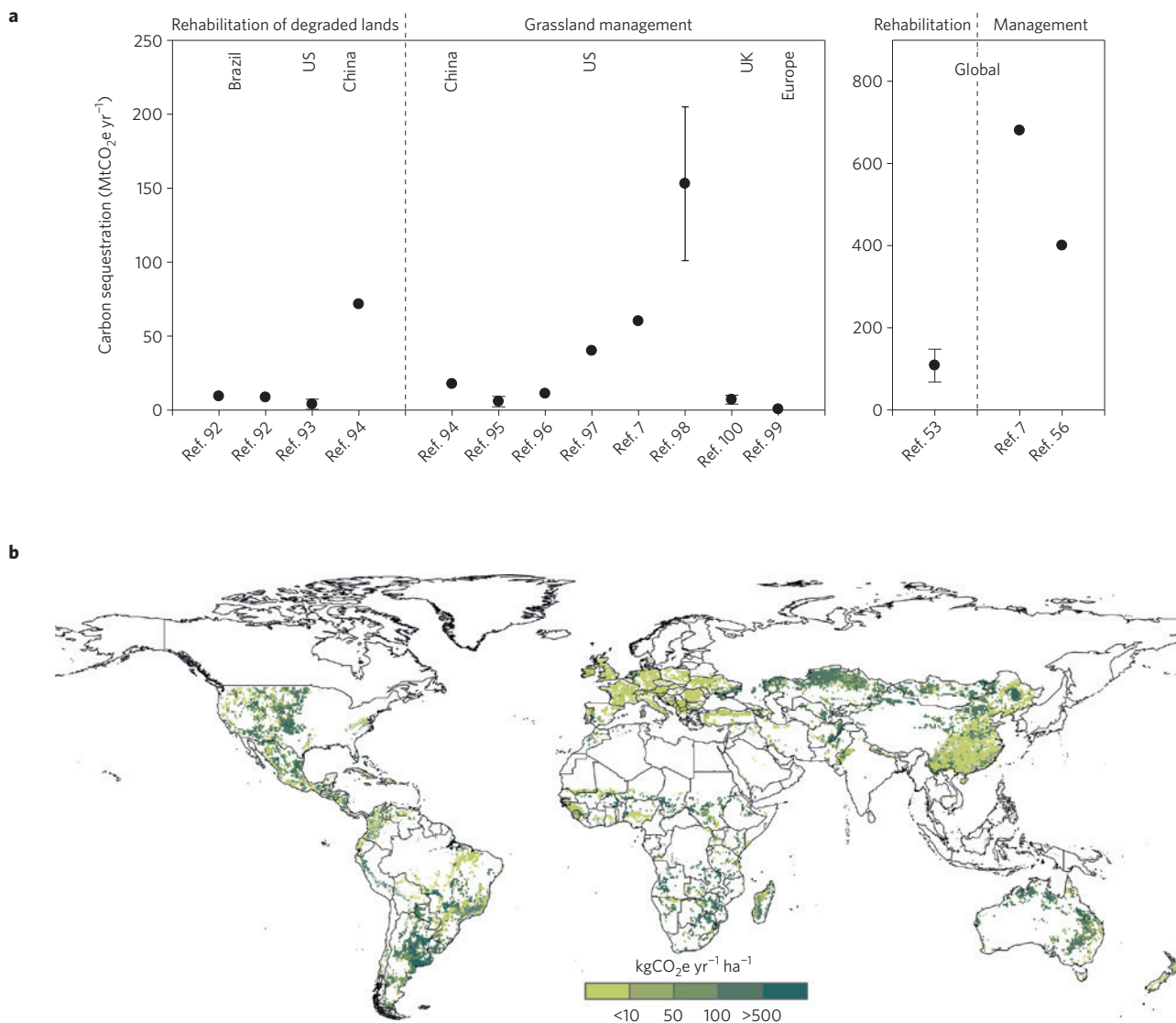
### Moderation of demand for livestock products

Projections of food demand that are based on growing populations and increasing per capita wealth suggest that demand for food will increase by 60–100% by 2050<sup>67</sup>. Given that the resource-use efficiency of livestock production is low in comparison with crops, and that about a third of the world's cereal production is fed to animals<sup>1</sup>, reducing consumption of livestock products in some places could greatly reduce the need for more food. On average, the production of beef protein requires about 50 times more land than the production of vegetable proteins<sup>68</sup>, and GHG emissions excluding land-use change are about 100 times higher. Although meat now represents only 15% of the total energy in the global human diet, approximately 80% of agricultural land is used for animal grazing or the production of feed and fodder for animals<sup>1</sup>, but this does include extensive grasslands in areas where other forms of agriculture would be extremely challenging.

Increases in the demand for meat have been used in a wide range of scenario analyses of the agricultural sector<sup>41,69</sup> (see the Supplementary Information for a description of the scenarios tested). By contrast, with global adoption of healthy diets (following healthy eating recommendations<sup>70</sup>), adequate food production in 2050 could be achieved on less agricultural land than is used at present, allowing the regrowth of natural vegetation and resulting in a reduction of GHG emissions of 4.3 GtCO<sub>2</sub>e yr<sup>-1</sup> relative to the baseline. More extreme scenarios would yield emissions reductions of 5.8, 6.4 and 7.8 GtCO<sub>2</sub>e yr<sup>-1</sup>, respectively, for no-ruminant meat, no-meat and no-animal product scenarios. In addition to reducing pressure on agricultural land, a global transition to a healthy, low-meat diet would reduce the mitigation costs, particularly in the energy sector, of achieving a 450 ppm CO<sub>2</sub>e stabilization target in 2050<sup>71</sup>. Reducing the demand for livestock products is, at least in theory, a powerful mitigation option<sup>11,72–75</sup>.

Reduced consumption could have substantial beneficial effects largely through its ability to create 'spare land' that can be used for either bioenergy or C sequestration by afforestation<sup>7,76</sup>. Switching to a low-animal-product diet that converges on the global average energy demand in the year 2000 (that is, 2,800 kcal per capita per day, compared with the global mean of 3,100 kcal per capita per day in the reference case) gave emissions reductions of 0.7–7.3 GtCO<sub>2</sub>e yr<sup>-1</sup>, if the spared land is used for bioenergy, and 4.6 GtCO<sub>2</sub>e yr<sup>-1</sup> if afforestation is assumed.

Dietary changes might occur in various ways, such as through changes in consumer preferences or via stringent climate policies and emissions pricing in the food and agriculture sectors that change the relative prices in favour of low-emissions food and thereby cause a shift in consumption<sup>77,78</sup>. However, emissions pricing could push up global food prices and reduce consumption in low-income regions, with negative impacts on food security<sup>79</sup>. Such impacts on global prices could be avoided, however, by implementing emissions pricing on food consumed rather than on food produced<sup>77</sup>.



**Figure 4 | The carbon sequestration potential of grazing lands. a**, The mitigation potentials for carbon sequestration in grasslands through rangeland rehabilitation and grazing management are shown for selected regions (left) and the globe as a whole (right). Data are from the references indicated. **b**, The spatial distribution of the carbon sequestration potential in rangelands<sup>56</sup>.

For demand-side options, determining a technical potential is not useful as it would be based on a rather implausible ‘no animal products’ scenario. Likewise, holding back consumption in the already low-meat diets of developing regions may be regarded as undesirable. With regard to plausible low-meat diets, we find a mitigation potential of 4.3–6.4 GtCO<sub>2</sub>e yr<sup>-1</sup>, with about 1–2 GtCO<sub>2</sub>e yr<sup>-1</sup> coming from process emissions, mostly CH<sub>4</sub> and N<sub>2</sub>O, and the remainder from land-use change CO<sub>2</sub> (see Supplementary Information). As assumptions on bioenergy grown on spare land would introduce another important source of uncertainty, we focus — where relevant — on afforestation of the abandoned land. In addition, we do not address waste reduction in the demand-side options, as this has mostly been analysed for the entire food system rather than for livestock specifically.

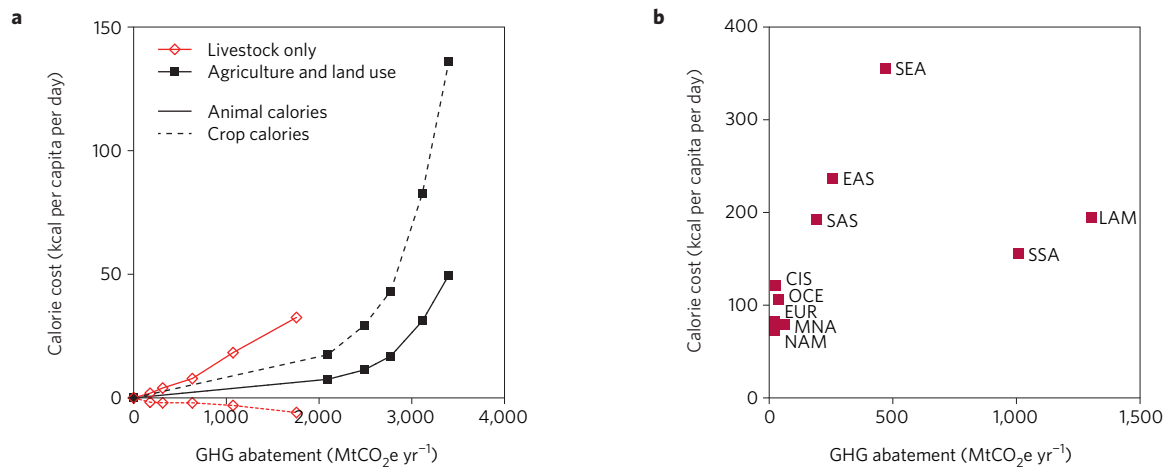
**Mitigation potentials, market effects and trade-offs**

The success of the practices described above in mitigating GHG from the livestock sector is largely dependent on their adoption. This in turn depends on public and private incentives, public policies and taxes, costs and logistics of implementation, and trade-offs between practices and with other sectors. Some of these

critical factors and their ability to yield economically feasible mitigation alternatives are described below.

**Economic supply-side mitigation potentials.** Although the technical mitigation potential for livestock is substantial, the share that can be achieved at reasonable economic cost is likely to be much smaller. Our yardstick is a marginal abatement cost of up to US\$50 per tCO<sub>2</sub>e, which is higher than current carbon market prices and those of the recent past, but still lower than the carbon price needed to make significant impacts on global GHG emissions across all sectors<sup>80</sup>.

The 2030 mitigation potentials for animal GHG emissions at unit costs of US\$20, US\$50 and US\$100 per tCO<sub>2</sub>e were estimated to be 175, 200 and 225 MtCO<sub>2</sub>e yr<sup>-1</sup>, respectively<sup>81</sup>. For measures targeting soil carbon sequestration in grazing lands, higher mitigation levels of 250, 375 and 750 MtCO<sub>2</sub>e yr<sup>-1</sup> by 2030 were estimated at these prices. Another global assessment estimated that animal GHG emissions could be reduced by 136, 193 and 228 MtCO<sub>2</sub>e yr<sup>-1</sup> by 2030 at unit costs of US\$20, US\$50 and US\$100 per tCO<sub>2</sub>e (ref. 27). The economic mitigation potentials in these studies were calculated without considering market



**Figure 5 | Total abatement calorie cost (TACC) of mitigation policies.** **a**, Policies targeting either the livestock sector only or all agricultural and land-use change GHG emissions with a carbon tax from US\$0–100 per tCO<sub>2</sub>e and decomposed by calorie source. The six plotted points (squares and diamonds) on each line correspond to the different levels of tax at US\$0, 5, 10, 20, 50 and 100. **b**, The TACC values of a mitigation policy that targets both agricultural and land-use change emissions with a carbon tax of US\$100 per tCO<sub>2</sub>e are shown by region (calculations based on ref. 79). NAM, North America; MNA, Middle East and North Africa; EUR, Europe; OCE, Oceania; CIS, Commonwealth of Independent States; SAS, South Asia; EAS, East Asia; SEA, Southeast Asia; SSA, Sub-Saharan Africa; LAM, Latin America.

interactions and their effects on reallocating livestock production, trade, land use and nutrition in the presence of mitigation policies. Some important consequences of these market interactions are discussed below.

**Emissions leakage.** This occurs when the mitigation policies used to reduce livestock emissions in one region cause production to fall, increasing the imports of livestock commodities to that region, thereby raising production and associated emissions in the exporting regions. This market-based migration of production can potentially lower the efficacy of mitigation policies, but if policies rely on positive incentives such as mitigation subsidies rather than negative incentives such as a carbon tax, it may be possible to reduce emissions without lowering production and thereby prevent leakage. If negative incentives are used, leakage can only be eliminated if the incentives are applied to all global livestock emissions. Annual reductions in livestock emissions of 163 MtCO<sub>2</sub>e yr<sup>-1</sup> have been estimated in response to a US\$27 per tCO<sub>2</sub>e carbon tax on agricultural emissions in industrialized (Annex I) countries<sup>82</sup>. However, 35% of this reduction in emissions is estimated to be offset by increased emissions in developing (non-Annex I) countries. A sensitivity analysis of the trade elasticities, which are critical determinants of the leakage rate, placed the mean leakage rate (with a 95% confidence) between 16% and 56%.

Two thirds of the emissions reduction achieved by a tax on livestock CH<sub>4</sub> emissions in industrialized (Annex I) countries could be leaked via increased emissions in developing countries<sup>83</sup>. Emissions leakage rates of 22–91% have been estimated for different economic scenarios in the EU<sup>23,84</sup>.

Emissions leakage could also be a problem for demand-side mitigation policies that seek to reduce the consumption of livestock products in some countries. Farmers in these countries may try to export their surplus production at lower prices to minimize their losses, increasing consumption in other regions, at least over the short term. As with supply-side mitigation policies, emissions leakage can also limit the effectiveness of consumption-side policies. The magnitude of this response greatly depends on the flexibility of trade and the sensitivity of consumers to price changes, among other things<sup>77,82</sup>. These findings highlight the importance of coordinated global mitigation policies to control for emissions leakages.

**Rebound effects.** The responses of economic agents to new technologies and practices that lower production costs and output prices are known as rebound effects<sup>80,84,85</sup>. This is important for policies that aim to mitigate land-use change emissions by intensifying production, as lower prices will drive higher consumption levels. If developing regions improved livestock feed efficiency by 25%, 371 MtCO<sub>2</sub>e yr<sup>-1</sup> could be avoided, but those savings could be up to three times larger if the demand for livestock products was held constant<sup>66</sup>. Similar rebound effects have been found for livestock efficiency improvements and waste reduction using different models<sup>86</sup>. In contrast, Cohn *et al.*<sup>87</sup> showed that policies fostering grassland intensification in Brazil were not significantly affected by rebound effects, and that global GHG savings from unilateral mitigation efforts by Brazil are only marginally affected. This demonstrates the importance of considering the attenuating influence of market interactions between economic agents and regions when assessing the effectiveness of mitigation policies.

**Emissions pricing and demand-side potentials.** Total abatement calorie cost (TACC) curves have been proposed as a measure of the change in food availability in response to emissions reductions from a GHG tax<sup>79</sup>. Figure 5a shows the strong relationship between GHG mitigation in the livestock sector and the availability of calories of animal origin. The overall abatement potential (considering all animal and land-use change emissions sources) under such a policy would be 1.8 GtCO<sub>2</sub>e yr<sup>-1</sup> with a demand reduction of about 27 kcal per capita per day at US\$100 per tCO<sub>2</sub>e. This relatively minor calorie cost is due to the overall small share of calories of animal origin in human diets: 18% globally. When implementing the GHG tax on all agricultural non-CO<sub>2</sub> emissions and CO<sub>2</sub> emissions from land-use change, the total abatement potential at US\$100 per tCO<sub>2</sub>e is about 3.4 GtCO<sub>2</sub>e yr<sup>-1</sup> and the corresponding calorie cost is about 190 kcal per capita per day. The calorie cost is low at taxes below US\$20 per tCO<sub>2</sub>e and most of the mitigation comes from avoided deforestation. Above this carbon price, further mitigation comes mostly from the agricultural sector, with significant impacts on food availability. The relationship between climate mitigation and food availability varies substantially across regions (Fig. 5b), which is a crucial consideration for targeting mitigation efforts.

Potential shifts in meat and dairy demand from the implementation of a GHG tax on food in the EU have been estimated<sup>77</sup>. Owing

to the relatively inelastic demand for meat and dairy products, the direct mitigation potential from GHG pricing of food was found to be small. For a climate tax on animal food equivalent to €60 per tCO<sub>2</sub>e, the estimated emissions reduction was about 32 MtCO<sub>2</sub>e yr<sup>-1</sup> — mainly from shifting from beef to other meats — which is about 7% of current GHG emissions in EU agriculture. However, this structural change in demand away from beef would also reduce total land requirements for food. If that spare land were used for bioenergy that is used as a substitute for fossil fuel energy, the mitigation effect of this policy intervention could be several times larger.

## Discussion

The mitigation potential of the livestock sector could represent up to 50% of the global mitigation potential of the agriculture, forestry and land-use sector. This is significant, but most of this potential has yet to be realized, due to low adoption rates of technical practices and uncertainties and trade-offs associated with attempts to reduce the consumption of livestock products. At the same time, establishing the societal impacts of land-sparing opportunities, in terms of livelihoods, employment, economics, gender and equity, is essential for understanding their feasibility. This area needs to receive urgent attention due to its policy relevance.

There is also a need to increase investment in the livestock sector in the developing world so that it becomes more market-oriented<sup>88</sup>. This could catalyse the adoption of practices for sustainably intensifying the sector in addition to mitigating emissions. Understanding the interactions at global and local scales between mitigation and adaptation in livestock systems will also be essential for identifying the practices that create the largest synergies and reduce the trade-offs between these goals<sup>89</sup>.

The continuing trend of increasing global consumption of meat is not compatible with reducing GHG emissions from agriculture. There is a need for research to understand what types of knowledge or interventions could limit the global growth in livestock consumption without threatening the food security or nutritional outcomes of people in developing countries. Public health policies offer one opportunity for moderating the overconsumption of livestock products by the growing middle classes that would be likely to have beneficial effects on both health and climate change mitigation<sup>73</sup>. On the other hand, the world food system has never had to react to planned, voluntary reductions in food consumption, so very few policy alternatives to reduce consumption equitably have been designed or tested<sup>90</sup>.

Limiting the rise in emissions from the livestock sector is challenging. There are opportunities for simultaneously increasing productivity and decreasing emissions intensity, but these run the risk of successful farmers expanding production, thus limiting the benefits in terms of total emissions. Thus addressing leakage and rebound impacts with the livestock sectors, and between the livestock and other sectors, will be essential. Reducing the global consumption of livestock products would bring considerable benefits in terms of lower agricultural emissions, but much more research is needed on how this could be achieved without negative trade-offs in some parts of the livestock sector. Reducing the large and growing contribution of the livestock sector to climate change while also ensuring that global nutritional security and health needs are supported is an urgent global research and investment priority.

Received 12 November 2015; accepted 4 January 2016;  
published online 21 March 2016

## References

1. *Livestock's Long Shadow: Environmental Issues and Options* (FAO, 2006).
2. Thornton, P. K. Livestock production: recent trends, future prospects. *Phil. Trans. R. Soc. B* **365**, 2853–2867 (2010).
3. Herrero, M., Thornton, P. K., Gerber, P. & Reid, R. S. Livestock, livelihoods and the environment: understanding the trade-offs. *Curr. Opin. Environ. Sust.* **1**, 111–120 (2009).
4. *The State of Food and Agriculture: Livestock in the Balance* (FAO, 2009).
5. Rosegrant, M. W. *et al.* in *Agriculture at a Crossroads* (eds McIntyre, B. D. *et al.*) 307–376 (Island, 2009).
6. Capper, J. L., Cady, R. A. & Bauman, D. E. The environmental impact of dairy production: 1944 compared with 2007. *J. Anim. Sci.* **87**, 2160–2167 (2009).
7. Smith, P. *et al.* How much land based greenhouse gas mitigation can be achieved without compromising food security and environmental goals? *Glob. Change Biol.* **19**, 2285–2302 (2013).
8. IPCC 2006 *IPCC Guidelines for National Greenhouse Gas Inventories* Vol. 4 (eds Eggleston, S. *et al.*) Ch. 10 (Cambridge Univ. Press, 2006).
9. *Emission Database for Global Atmospheric Research (EDGAR) v.4.2* (JRC and PBL, accessed 21 September 2014); <http://edgar.jrc.ec.europa.eu>
10. *Global Anthropogenic Non-CO<sub>2</sub> Greenhouse Gas Emissions: 1990–2020* (US EPA, 2006).
11. Popp, A., Lotze-Campen, H. & Bodirsky, B. Food consumption, diet shifts and associated non-CO<sub>2</sub> greenhouse gases from agricultural production. *Glob. Environ. Change* **20**, 451–462 (2010).
12. Tubiello, F. N. *et al.* The FAOSTAT database of greenhouse gas emissions from agriculture. *Environ. Res. Lett.* **8**, 015009 (2013).
13. Herrero, M. *et al.* Biomass use, production, feed efficiencies, and greenhouse gas emissions from global livestock systems. *Proc. Natl Acad. Sci. USA* **110**, 20888–20893 (2013).
14. *Greenhouse Gas Emissions From the Dairy Sector: A Life Cycle Assessment* (FAO, 2010).
15. Gerber, P. *et al.* *Tackling Climate Change Through Livestock—A Global Assessment of Emissions and Mitigation Opportunities* (FAO, 2013).
16. Bodirsky, B. L. *et al.* N<sub>2</sub>O emissions from the global agricultural nitrogen cycle — current state and future scenarios. *Biogeosciences* **9**, 4169–4197 (2012).
17. *Greenhouse Gas Emissions from Ruminant Supply Chains—A Global Life Cycle Assessment* (FAO, 2013).
18. *Greenhouse Gas Emissions from Pig and Chicken Supply Chains—A Global Life Cycle Assessment* (FAO, 2013).
19. Fowler, D. *et al.* Atmospheric composition change: ecosystems–atmosphere interactions. *Atmos. Environ.* **43**, 5193–5267 (2009).
20. Davidson, E. A. The contribution of manure and fertilizer nitrogen to atmospheric nitrous oxide since 1860. *Nature Geosci.* **2**, 659–662 (2009).
21. Reay, D. S. *et al.* Global agriculture and nitrous oxide emissions. *Nature Clim. Change* **2**, 410–416 (2012).
22. Rypdal, K. & Winiwarter, W. Uncertainties in greenhouse gas emission inventories, evaluation, comparability and implications. *Environ. Sci. Policy* **4**, 107–116 (2001).
23. Leip, A. *et al.* *Evaluation of the Livestock Sector's Contribution to the EU Greenhouse Gas Emissions (GGELS)* (Joint Research Centre, European Commission, 2010).
24. Monni, S., Perala, P. & Regina, K. Uncertainty in agricultural CH<sub>4</sub> and N<sub>2</sub>O emissions from Finland — possibilities to increase accuracy in emission estimates. *Mitig. Adapt. Strat. Glob. Change* **12**, 545–571 (2007).
25. Weiss, F. & Leip, A. Greenhouse gas emissions from the EU livestock sector: a life cycle assessment carried out with the CAPRI model. *Agric. Ecosyst. Environ.* **149**, 124–134 (2012).
26. Leip, A. *et al.* Impacts of European livestock production: nitrogen, sulphur, phosphorus and greenhouse gas emissions, land use, water eutrophication and biodiversity. *Environ. Res. Lett.* **10**, 115004 (2015).
27. *Global Mitigation of Non-CO<sub>2</sub> Greenhouse Gases: 2010–2030* (US EPA, 2013).
28. Westhoek, H. *et al.* *The Protein Puzzle* (PBL, 2011).
29. Havlik, P. *et al.* Crop productivity and the global livestock sector: Implications for land use change and greenhouse gas emissions. *Am. J. Agric. Econ.* **95**, 442–448 (2013).
30. de Vries, M. and de Boer, I. J. M. Comparing environmental impacts for livestock products: a review of life cycle assessments. *Livest. Sci.* **128**, 1–11 (2009).
31. Cederberg, C., Hedenus, F., Wirsenius, S. & Sonesson, U. Trends in greenhouse gas emissions from consumption and production of animal food products: implications for long-term climate targets. *Animal* **7**, 330–340 (2013).
32. Hristov, A. N. *et al.* *Mitigation of Greenhouse Gas Emissions in Livestock Production. A Review of Technical Options for Non-CO<sub>2</sub> Emissions* (FAO, 2013).
33. Boadi, D., Benchaar, C., Chiquette, J. & Massé, D. Mitigation strategies to reduce enteric methane emissions from dairy cows: update review. *Can. J. Anim. Sci.* **84**, 319–335 (2004).
34. Martin, C., Morgavi, D. P. & Doreau, M. Methane mitigation in ruminants: from microbe to the farm scale. *Animal* **4**, 351–365 (2010).
35. Cottle, D. J., Nolan, J. V. & Wiedemann, S. G. Ruminant enteric methane mitigation: a review. *Anim. Prod. Sci.* **51**, 491–514 (2011).



36. Hristov, A. N. *et al.* An inhibitor persistently decreased enteric methane emission from dairy cows with no negative effect on milk production. *Proc. Natl Acad. Sci. USA* **112**, 10663–10668 (2015).
37. Thornton, P. K. & Herrero, M. The potential for reduced methane and carbon dioxide emissions from livestock and pasture management in the tropics. *Proc. Natl Acad. Sci. USA* **107**, 19667–19672 (2010).
38. Ripple, W. J. *et al.* Ruminants, climate change and climate policy. *Nature Clim. Change* **4**, 2–5 (2014).
39. Hristov, A. N. Historic, pre-European settlement, and present-day contribution of wild ruminants to enteric methane emissions in the United States. *J. Anim. Sci.* **90**, 1371–1375 (2012).
40. Walsh, B. *et al.* New feed sources key to ambitious climate targets. *Carbon Balance Manage.* **10**, 26 (2015).
41. Smith P. *et al.* Greenhouse gas mitigation in agriculture. *Phil. Trans. R. Soc. B* **363**, 789–813 (2008).
42. Chadwick, D. *et al.* Manure management: implications for greenhouse gas emissions. *Anim. Feed Sci. Technol.* **166–167**, 514–531 (2011).
43. Chadwick, D. Emissions of ammonia, nitrous oxide and methane from cattle manure heaps: effect of compaction and covering. *Atmos. Environ.* **39**, 87–799 (2005).
44. Thomsen, I. K., Pedersen, A. R., Nyord, T. & Petersen, S. O. Effects of slurry pre-treatment and application technique on short-term N<sub>2</sub>O emissions as determined by a new non-linear approach. *Agric. Ecosyst. Environ.* **136**, 227–235 (2010).
45. Clemens, J., Trimborn, M., Weiland, P. & Amon, B. Mitigation of greenhouse gas emissions by anaerobic digestion of cattle slurry. *Agric. Ecosyst. Environ.* **112**, 171–177 (2006).
46. van Groenigen J. W. *et al.* Nitrous oxide emissions from silage maize fields under different mineral nitrogen fertilizer and slurry applications. *Plant Soil* **263**, 101–111 (2004).
47. Webb, J., Pain, B., Bittman, S. & Morgan, J. The impacts of manure application methods on emissions of ammonia, nitrous oxide and on crop response—a review. *Agric. Ecosyst. Environ.* **137**, 39–46 (2010).
48. van Groenigen, J. W., Velthof, G. L., Oenema, O. & van Groenigen, K. J. Towards an agronomic assessment of N<sub>2</sub>O emissions: a case study for arable crops. *Eur. J. Soil Sci.* **61**, 903–913 (2010).
49. Smith, K. A. & Conen, F. Impacts of land management on fluxes of trace greenhouse gases. *Soil Use Manag.* **20**, 255–263 (2004).
50. Snyder, C. S., Bruulsema, T. W., Jensen, T. L. & Fixen, P. E. Review of greenhouse gas emissions from crop production systems and fertilizer management effects. *Agric. Ecosyst. Environ.* **133**, 247–266 (2009).
51. Clough T. J. *et al.* The mitigation potential of hippuric acid on N<sub>2</sub>O emissions from urine patches: an *in situ* determination of its effect. *Soil Biol. Biochem.* **41**, 2222–2229 (2009).
52. Betram J. E. *et al.* Hippuric acid and benzoic acid inhibition of urine derived N<sub>2</sub>O emissions from soil. *Glob. Change Biol.* **15**, 2067–2077 (2009).
53. Conant, R. T. & Paustian, K. Potential soil carbon sequestration in overgrazed grassland ecosystems. *Glob. Biogeochem. Cycles* **16**, 1143–1152 (2002).
54. Ojima, D. S. *et al.* Modeling the effects of climatic and CO<sub>2</sub> changes on grassland storage of soil C. *Water Air Soil Pollut.* **70**, 643–657 (1993).
55. Reid, R. S. *et al.* Is it possible to mitigate greenhouse gas emissions in pastoral ecosystems of the tropics? *Environ. Dev. Sust.* **6**, 91–109 (2004).
56. Henderson, B. *et al.* Greenhouse gas mitigation potential of the world's grazing lands: modelling soil carbon and nitrogen fluxes of mitigation practices. *Agric. Ecosyst. Environ.* **207**, 91–100 (2015).
57. DeFries, R. & Rosenzweig, C. Toward a whole-landscape approach for sustainable land use in the tropics. *Proc. Natl Acad. Sci. USA* **107**, 19627–19632 (2010).
58. Gill, M., Smith, P. & Wilkinson, J. M. Mitigating climate change: the role of domestic livestock. *Animal* **4**, 323–333 (2010).
59. Smith, P. Delivering food security without increasing pressure on land. *Glob. Food Sec.* **2**, 18–23 (2013).
60. Foresight *The Future of Food and Farming* (The Government Office for Science, 2011).
61. Godfray, H. C. J. *et al.* Food security: the challenge of feeding 9 billion people. *Science* **327**, 812–818 (2010).
62. *International Assessment of Agricultural Knowledge, Science and Technology for Development* (IAASTD, 2009); <http://go.nature.com/O1GjwV>
63. Foley, J. A. *et al.* Solutions for a cultivated planet. *Nature* **478**, 337–342 (2011).
64. Mueller, N. D. *et al.* Closing yield gaps through nutrient and water management. *Nature* **490**, 254–257 (2012).
65. Tilman, D., Balzer, C., Hill, J. & Befort, B. L. Global food demand and the sustainable intensification of agriculture. *Proc. Natl Acad. Sci. USA* **108**, 20260–20264 (2011).
66. Valin, H. *et al.* Agricultural productivity and GHG emissions in developing countries: what future trade-offs between mitigation and food security? *Environ. Res. Lett.* **8**, 035019 (2013).
67. *Reaping the Benefits: Science and the Sustainable Intensification of Global Agriculture* (Royal Society, 2009).
68. Nijdam, D., Rood, T. & Westhoek, H. The price of protein: review of land use and carbon footprints from life cycle assessments of animal food products and their substitutes. *Food Policy* **37**, 760–770 (2012).
69. Valin, H. *et al.* The future of food demand: understanding differences in global economic models. *Agric. Econ.* **45**, 51–67 (2014).
70. Willett, W. C. *Eat, Drink, and Be Healthy: the Harvard Medical School Guide to Healthy Eating* (Simon and Schuster, 2001).
71. Stehfest, E. *et al.* Climate benefits of changing diet. *Climatic Change* **95**, 83–102 (2009).
72. Bajželj, B. *et al.* The importance of food demand management for climate mitigation. *Nature Clim. Change* **4**, 924–929 (2014).
73. Tilman, D. & Clark, M. Global diets link environmental sustainability and human health. *Nature* **515**, 518–522 (2014).
74. Hedenus, F., Wirsenius, S. & Johansson, D. J. The importance of reduced meat and dairy consumption for meeting stringent climate change targets. *Climatic Change* **124**, 79–91 (2014).
75. Westhoek, H. *et al.* Food choices, health and environment: effects of cutting Europe's meat and dairy intake. *Glob. Environ. Change.* **26**, 126–135 (2014).
76. Erb, K.-H., Haberl, H. & Plutzer, C. Dependency of global primary bioenergy crop potentials in 2050 on food systems, yields, biodiversity conservation and political stability. *Energy Policy* **47**, 260–269 (2012).
77. Wirsenius, S., Hedenus, F. & Mohlin, K. Greenhouse gas taxes on animal food products: rationale, tax scheme and climate mitigation effects. *Climatic Change* **108**, 159–184 (2011).
78. Edjabou, L. D. & Smed, S. The effect of using consumption taxes on foods to promote climate friendly diets—the case of Denmark. *Food Policy* **39**, 84–96 (2013).
79. Havlik, P. *et al.* Climate change mitigation through livestock system transitions. *Proc. Natl Acad. Sci. USA* **111**, 3709–3714 (2014).
80. Edenhofer, O. *et al.* (eds) *Climate Change 2014: Mitigation of Climate Change* (IPCC, Cambridge Univ. Press, 2014).
81. Smith, P. *et al.* in *Climate Change 2007: Mitigation of Climate Change* (eds Metz, B. *et al.*) Ch. 8 (IPCC, Cambridge Univ. Press, 2007).
82. Golub, A. *et al.* Global climate policy impacts on livestock, land use, livelihoods, and food security. *Proc. Natl Acad. Sci. USA* **110**, 20894–20899 (2013).
83. Key, N. & Tallard, G. Mitigating methane emissions from livestock: a global analysis of sectoral policies. *Climatic Change*, **12**, 387–414 (2012).
84. Van Doorslaer, B. *et al.* *An Economic Assessment of GHG Mitigation Policy Options for EU Agriculture* (Joint Research Centre, European Commission, 2015).
85. Lambin, E. F. & Meyfroidt, P. Global land use change, economic globalization, and the looming land scarcity. *Proc. Natl Acad. Sci. USA* **108**, 3465–3472 (2011).
86. Stehfest, E., Berg, M. V. D., Woltjer, G., Msangi, S. & Westhoek, H. Options to reduce the environmental effects of livestock production - Comparison of two economic models. *Agric. Syst.* **114**, 38–53 (2013).
87. Cohn, A. *et al.* Cattle ranching intensification in Brazil can reduce global greenhouse gas emissions by sparing land from deforestation. *Proc. Natl Acad. Sci. USA* **111**, 7236–7241 (2014).
88. Herrero, M. *et al.* Smart investments in sustainable food production: revisiting mixed crop-livestock systems. *Science* **327**, 822–825 (2010).
89. Vervoort, J. *et al.* Challenges to scenario-guided adaptive action on food security under climate change. *Glob. Environ. Change* **28**, 383–394 (2014).
90. Garnett, T., Mathewson, S., Angelides, P. & Borthwick, F. *Policies and Actions to Shift Eating Patterns. What Works?* (Food and Climate Research Network, 2015).
91. Conant, R. T., Paustian, K. & Elliott, E. T. Grassland management and conversion into grassland: effects on soil carbon. *Ecol. Appl.* **11**, 343–355 (2001).
92. Carvalho, J. L. N. *et al.* Potential of soil carbon sequestration in different biomes of Brazil. *Rev. Bras. Cienc. Solo* **34**, 277–289 (2010).
93. Lal, R. Soil carbon sequestration impacts on global climate change and food security. *Science* **304**, 1623–1627 (2004).
94. Wang, S. *et al.* Management and land use change effects on soil carbon in northern China's grasslands: a synthesis. *Agric. Ecosyst. Environ.* **142**, 329–340 (2011).
95. Follett, R. F. & Schumann, G. E. in *Grassland: A Global Resource* (ed. McGilloway, D. A.) 264–277 (Wageningen, 2005).
96. Schuman, J. E., Janzen, H. H. & Herrick, J. E. Soil carbon sequestration and potential carbon sequestration in rangelands. *Environ. Pollut.* **116**, 391–396 (2002).
97. Morgan, J. *et al.* Carbon sequestration in agricultural lands of the United States. *J. Soil Water Cons.* **65**, 6–13 (2011).

98. Lal, R. Carbon sequestration in dryland ecosystems. *Environ. Manage.* **33**, 528–544, (2003).
99. Bellarby, J. *et al.* Livestock greenhouse gas emissions and mitigation potential in Europe. *Glob. Change Biol.* **19**, 3–18 (2013).
100. Fitton, N. *et al.* Greenhouse gas mitigation potential of agricultural land in Great Britain. *Soil Use Manage.* **27**, 491–501 (2011).

### Acknowledgements

This paper constitutes an output of the Belmont Forum/FACCE-JPI funded DEVIL project (NE/M021327/1). Financial support from the CGIAR Program on Climate Change, Agriculture and Food Security (CCAFS) and the EU-FP7 AnimalChange project is also recognized. P.K.T. acknowledges the support of a CSIRO McMaster Research Fellowship.

### Author contributions

M.H. conceived the study and prepared the manuscript. All authors analysed data and contributed to the writing and editing of the manuscript.

### Additional information

Supplementary information is available in the [online version of the paper](#). Reprints and permissions information is available online at [www.nature.com/reprints](http://www.nature.com/reprints). Correspondence should be addressed to M.H.

### Competing financial interests

The authors declare no competing financial interests.